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1 Short Title: Eradicating *Phytophthora* species by composting

2 **Potential for eradication of the exotic plant pathogens**

3 ***Phytophthora kernoviae* and *Phytophthora ramorum* during**
4 **composting**

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10 **Abstract**

11 *Keywords* *Phytophthora pseudosyringae* *Rhododendron ponticum* *Vaccinium myrtillus*

12 Temperature

13 Temperature and exposure time effects on *Phytophthora kernoviae* and *P. ramorum*
14 viability were examined in flasks of compost and in a large-scale composting system
15 containing plant waste. Cellophane, rhododendron leaf and peat-based inoculum of *P.*
16 *kernoviae* and *P. ramorum* isolates were used in flasks, and naturally infected leaves were
17 inserted into a large-scale system. Exposures of 5 and 10 days respectively at a mean
18 temperature of 35°C in flask and large-scale composts reduced *P. kernoviae* and *P.*
19 *ramorum* inocula to below detection limits using semi-selective culturing. Although *P.*
20 *ramorum* was undetectable after one-day exposure of inoculum to compost at 40°C in
21 flasks, it survived on leaves exposed to a mean temperature of 40.9°C for 5 days in a
22 large-scale composting system. No survival of *P. ramorum* was detected after exposure

1 of infected leaves for 5 days to a mean temperature of $\geq 41.9^{\circ}\text{C}$ (32.8°C for *P.*
2 *kernoviae*) or for 10 days at $\geq 31.8^{\circ}\text{C}$ (25.9°C for *P. pseudosyringae* on infected bilberry
3 stems) in large-scale systems. Fitted survival probabilities of *P. ramorum* on infected
4 leaves exposed in a large-scale system for 5 days at 45°C or for 10 days at 35°C were less
5 than 3%, for an average initial infection level of leaves of 59.2%. RNA quantification to
6 measure viability was shown to be unreliable in environments that favour RNA
7 preservation: high levels of ITS1 RNA were recovered from *P. kernoviae* and *P.*
8 *ramorum* infected leaves exposed to composting plant wastes at $>53^{\circ}\text{C}$, when all culture
9 results were negative.

10 **Introduction**

11 The oomycetes *Phytophthora kernoviae* and *P. ramorum* are introduced pathogens to the
12 United Kingdom (UK). *P. ramorum* has also been introduced to many other European
13 countries, as well as the USA where it causes the widespread death of native oaks in
14 California and Oregon known as sudden oak death (Harnik et al., 2004; Jennings, 2008).
15 More recently in the UK, the widespread death of Japanese larch (*Larix kaempferi*)
16 caused by *P. ramorum*, has been reported (Webber et al., 2010). *P. kernoviae* is also
17 present in New Zealand where it was first reported in 2006 (Sansford, 2008). Both
18 pathogens have a wide plant host range, including members of the *Fagaceae* such as
19 oaks and beech, members of the *Lauraceae*, as well as plants in several other taxonomic
20 families (Linderman & Davis, 2006). In the UK, *Rhododendron* has been particularly
21 susceptible (Sansford & Woodhall, 2007; Sansford, 2008) and the introduced and
22 invasive *Rhododendron ponticum* is an important intermediate host and source of
23 multiplication and promotion of pathogen inoculum for susceptible native plants such as

1 bilberry (*Vaccinium myrtillus*) (Beales et al., 2009) as well as for trees. Detached leaves
2 and stems infected with *P. ramorum* can remain infectious for several weeks and *P.*
3 *ramorum* and *P. kernoviae* can survive for at least two years in soil so that management
4 of plant wastes in infected areas is problematic, and poses an on-going phytosanitary risk
5 (Swain et al., 2006; Sansford & Woodhall, 2007; Sansford, 2008). Composting of
6 contaminated plant waste is a potential management option. Temperatures achieved in
7 well-managed composting systems are usually sufficient to ‘eradicate’ most plant
8 pathogens including *Phytophthora* species, where eradication is defined as ‘reducing the
9 inoculum to below a detectable limit’ (Noble & Roberts, 2004; Noble et al., 2009).
10 Exposing infected plant material for two weeks in composts reaching at least 55°C has
11 been shown to be effective in eliminating *P. ramorum* (Swain et al., 2006), although the
12 efficacy of shorter and cooler composting processes in achieving eradication has not been
13 investigated. Previous research into the effect of exposure of *P. ramorum* inoculum to
14 different time-temperature combinations *in vitro* and *in planta* has produced conflicting
15 results (Noble et al., 2009). This may at least partly be due to the inocula containing
16 different structures with differing resiliency such as chlamydospores, sporangia, oospores
17 and mycelial fragments. Jennings (2008) showed that *P. kernoviae* was eradicated *in vitro*
18 and *in planta* after exposure to temperature treatments that were effective in killing *P.*
19 *ramorum*. However, there are no previous studies on the efficacy of composting in
20 eradicating *P. kernoviae*.

21 Semi-selective culturing followed by morphological identification of *Phytophthora*
22 colonies is the usual method for detecting *Phytophthora* viability, but this does not
23 provide quantitative measures of viability of inocula (Noble et al., 2009). Swain et al.

(2006) used a PCR assay to detect *P. ramorum* DNA post-composting, but could not distinguish live and dead material. RNA quantification by real-time polymerase chain reaction (RT-PCR) has been used as an alternative method to measure viability of *Phytophthora* inoculum (Chimento et al., 2008). As part of this study, we have assessed this approach on composted samples with comparison to viability results obtained from semi-selective culture.

The aims of this work were to establish the minimum compost temperatures and exposure times required to eradicate *P. ramorum* and *P. kernoviae*. The influence of controlled temperatures on inocula containing different types of propagules of known isolates was examined in flasks of compost, followed by studying the influence of measured temperatures on naturally infected leaves in a large-scale composting system containing shredded plant wastes. The influences of compost temperatures and exposure times on *P. kernoviae* and *P. ramorum* were compared with the influences on another *Phytophthora* species that infects heathland bilberry, *Phytophthora pseudosyringae* (Beales et al., 2010).

Materials and Methods

***Phytophthora* species and inoculum**

P. kernoviae isolate CC95 (Food and Environment Research Agency [Fera], York) and *P. ramorum* isolate CC47 (Fera) were used for the bench-scale flask experiments. Each isolate was grown on *Rhododendron cuneatum* ‘Cunningham’s White’ leaves; cellophane discs by placing the discs over agar cultures; and in a peat-based medium (Noble et al., 2011). The *P. ramorum* leaf and cellophane inocula contained hyphae, sporangia, and chlamydospores; the *P. kernoviae* leaf and cellophane inocula contained hyphae,

1 sporangia, and oospores. Previous measurements conducted on the same types of leaf
2 inocula produced under the same conditions have shown a mean of eight sporangia and
3 four chlamydospores per 100 mm² for *P. ramorum* and five sporangia per 100 mm² for
4 *P. kernoviae*. The cellophane disc inoculum contained hyphae and sporangia at
5 concentrations of 180 colony forming units (cfu) per mm² for *P. ramorum* and 170 cfu
6 per mm² for *P. kernoviae*. The initial peat-based inoculum contained hyphae and
7 chlamydospores at a concentration of 2.3×10^6 cfu per g for *P. ramorum* and hyphae and
8 oospores at a concentration of 7.2×10^5 cfu per g for *P. kernoviae*. Three squares (10 × 10
9 mm) of leaf and cellophane inocula were placed individually in polyester mesh bags with
10 3 g of the composted green waste used in the flasks. Sample bags were also filled with 3
11 g peat inoculum.

12 For large-scale composting systems, naturally putative diseased leaves,
13 predominantly rhododendron and *Viburnum* spp., with symptoms typical of *P. kernoviae*
14 or *P. ramorum* and diseased bilberry stems with symptoms typical of *P. pseudosyringae*
15 were used. The levels of infection with each *Phytophthora* species in the samples of
16 putative diseased leaves or stems was determined as *post hoc* using semi-selective plate
17 culturing.

18 **RNA quantification**

19 Inoculum samples from the composting tests described below were analysed by reverse
20 transcriptase (RT) real-time PCR (RT-Taqman) assays, to quantify RNA, and hence live
21 *Phytophthora* cells, since RNA is very labile in exogenous conditions. RNA was
22 extracted from 100 mg of inoculum using the Qiagen RNeasy Plant Mini kit (part 74904).
23 DNA was removed according to the manufacturer's instructions by prolonged incubation

with RNase-free DNaseI. Two Taqman assays were designed to specifically detect *P. ramorum* and *P. kernoviae*, based on the assay of Hughes et al. (2006), which targeted the ITS1 region of nuclear ribosomal DNA. The *CoxI* region used previously (Chimento et al., 2008) was found not to be sufficiently polymorphic to allow design of assays specific to *P. ramorum* and *P. kernoviae*. Oligonucleotides were as follows (5'-3'): *P. ramorum*, forward = CTGGCTTCGGCTGGCT, reverse = AAAAGTGGGCTACTAGCTCAGAC, probe = 6'FAM-CTCTATCATGGCGAGCGCTTGAGCCT-TAMRA; *P. kernoviae*, forward = CTCGTTGGCAGTTTCGACTGT, reverse = GGTTTAAAAAAGAACTCTCAATCCGACT, probe = 6'FAM-AGTTCTATCAAACGATCGATTTGGGCTGCA-TAMRA. RT-Taqman reactions were run as follows in 25 µL volumes in an ABI7900 SDS machine: 1 × Absolute QPCR ROX Mix (Thermo, AB-1139); 5 U Superscript III reverse transcriptase (Invitrogen), 5 mM Dithiothreitol, 0.5 µM each oligonucleotide. Run parameters were: 42°C, 30'; 95°C, 10'; and 40 cycles of 95°C, 15" and 60°C, 1'. Cycle threshold (Ct) values were recorded and equivalent *Phytophthora* genomes per reaction were calculated using previous calibrations of known numbers of *P. ramorum* chlamydospores (using average 12 genomes each) and *P. kernoviae*, oospores (data not shown). These calibrations were only intended to provide a relative scale between composting conditions and not to indicate absolute numbers of spores or genomes.

Bench-scale flask composting

Composting was conducted in bench-scale equipment consisting of 'Quickfit' multiadapter flasks (Fisher Scientific) immersed in thermostatically controlled water baths, each holding up to four 2-L flasks (Noble et al., 2011). Each flask contained 1200

g of compost and was connected to ancillary equipment to aerate the compost for 2 min in every 15 min at a flow rate of 150 mL min⁻¹. The temperature of the compost was monitored with Squirrel[®] multipoint temperature data loggers (Grant Instruments) and gaseous conditions in the compost using methods described in Noble et al. (2011). The green waste compost used in the flasks was prepared in turned windrows using wastes from parks and gardens and fruit and vegetables for 9 months (Organic Recycling Ltd). Water was added to the compost to obtain an initial moisture content of 41 (±1) % w/w.

The following compost temperature × exposure time treatments were examined on three types of inoculum of *P. kernoviae* and *P. ramorum* in an incomplete factorial design experiment: 18 and 35°C for 1, 3, 5 and 10 days; 40 and 45°C for 1 and 3 days. There were between two and six replicate flasks of each temperature treatment (Table 1). Each flask contained sufficient sample bags of each *Phytophthora* species and inoculum type to enable a bag of each species and type to be retrieved at each of the exposure times.

Sample bags containing *P. kernoviae* and *P. ramorum* leaf inoculum exposed to compost at 35 and 45°C for 1 or 3 days were also kept for a further 14 days at 20°C in the compost to detect viability after potential constitutive or heat-induced ‘dormancy’. Two replicate flasks of each temperature treatment contained additional cellophane, leaf and peat inoculum of *P. ramorum*, which was tested for viability post-composting by RNA quantification. All the remaining *Phytophthora* inocula retrieved from sample bags post-composting were tested for viability by semi-selective plate culture on PDA + PARPH agar (Mitchell et al., 1986). Plates from cellophane and leaf inocula were classed as positive if mycelia or spores characteristic of *Phytophthora* were present; plates from peat samples were classed positive if germinating chlamydospores or oospores were

1 detected by the emergence of hyphae. The detection limit in the peat-based inoculum was
2 33 cfu per g for both *P. kernoviae* and *P. ramorum*.

3 **Large-scale composting**

4 Composts were prepared at Trengwainton Gardens, Penzance, Cornwall in open bays
5 measuring $2.9 \times 2.2 \times 2$ (height) m internally, and at Cannock Chase, Staffordshire in an
6 insulated cargo container measuring $3.7 \times 2.4 \times 2.6$ (height) m internally. The bays were
7 constructed of insulated wooden walls on three sides mounted on a concrete base. After
8 filling with c. 2.9 t wetted plant wastes (mainly rhododendron at Trengwainton and
9 bilberry at Cannock Chase) to a height of c. 1.5 m, the open ends of the bays or
10 containers were closed by horizontal wooden boards fitted into vertical slots on the ends
11 of the side walls, and the wastes were covered with an insulating waterproof layer. The
12 temperature of the wastes was allowed to rise through the natural composting process for
13 about seven days. Samples of putative *P. kernoviae* and *P. ramorum* infected leaves or *P.*
14 *pseudosyringae* infected bilberry stems were enclosed in 500×300 mm nylon mesh bags
15 containing 2 kg of the same plant wastes that was in the composting system, together
16 with a temperature probe connected to a data logger (Grant Instruments). The bags were
17 inserted into different locations of the composting systems (centre, corners, sides), at the
18 surface and at depths of 0.7 and 1.5 m (base), to obtain a range in exposure temperatures.
19 Gaseous conditions in the composts were measured as described previously.

20 Four compost batches were prepared in bays at Trengwainton from shredded and
21 wetted woody wastes, predominantly rhododendron (Table 2). In the first batch, 6–8
22 putative *P. kernoviae* or *P. ramorum* infected leaves were placed in each bag and 30
23 leaves were used as day 0 controls. In the subsequent three batches, 9 putative infected

1 leaves were inserted in the bags and 81–100 leaves were used as day 0 controls.
2 Depending on the batch, between 8 and 21 bags were inserted into the composting
3 wastes, and between 4 and 17 bags were retrieved after 5 and 10 days. In the third and
4 fourth replicate batches, bags of leaves were only positioned in the corners and sides of
5 the bay because these were considered to be the most critical positions in terms of
6 compost temperature and pathogen survival.

7 Three compost batches were prepared in the cargo container at Cannock Chase
8 from chopped and wetted bilberry waste (Table 2). In each batch, five putative *P.*
9 *pseudosyringae* infected bilberry stems were inserted in each of eight bags; 18–24
10 putative infected stems were used as day 0 controls.

11 After retrieval from the composts, putative *P. kernoviae* and *P. ramorum* infected
12 leaves were divided in two and one half used for semi-selective plate culturing and the
13 other half used for RNA analysis as described earlier. Viability of *P. pseudosyringae* on
14 retrieved stems was determined by semi-selective plate culturing.

15 **Statistical analysis**

16 Relationships between compost temperature and exposure time and *Phytophthora*
17 survival in the large-scale systems were investigated using a logistic regression (a
18 Generalized Linear Model with a logit link function). This compared the proportion of
19 the leaves or bilberry stems in each bag that were still positive for *P. kernoviae*, *P.*
20 *ramorum* or *P. pseudosyringae* at the end of the experiment. However, in order to
21 account for the fact not all leaves or stems were positive at the start of the experiment and
22 that this proportion differed between the compost batches, an offset variable (the
23 logarithm of the proportion of positive leaves or stems at the start of each compost batch)

1 was used. This allowed for the fact that bags from composts where larger proportions of
2 *Phytophthora* positives were observed at day 0 could be expected to have a greater
3 probability of containing *Phytophthora* positives at the end of the compost batch than
4 bags from other batches. All bags from the fourth compost batch from Trengwainton
5 were only used in statistical analysis for *P. ramorum* because none of the day 0 or post-
6 composting leaves were positive for *P. kernoviae*. All analyses were conducted in
7 Genstat 12.1 or 13.1.

8

9 **Results**

10 **Bench-scale flask composting**

11 During flask incubation, aerobic conditions were maintained in the composts, there were
12 low but detectable levels of ammonia, and the moisture content of the compost declined
13 slightly (Table 1).

14 The shift from positive to negative results was sudden and occurred in the same
15 temperature \times time treatment for almost all flask replicates, which prevented any
16 mathematical model to be fitted to the data (Table 2). In peat-based inoculum bags where
17 survival was detected post-composting, there were at least 889 cfu per g for *P. kernoviae*
18 and 367 cfu per g for *P. ramorum* at the end of the composting process, although the
19 mean values of the smallest numbers of detected propagules were slightly higher (Table
20 2). These values represent log reductions of 3.4 and 3.3 from the original inoculum
21 concentrations. There was consistent survival of the inoculum of both species in compost
22 at 18°C for up to 10 days; the exception was for *P. kernoviae* as cellophane inoculum,
23 which lost some viability after 1 and 3 days and did not survive after 5 and 10 days

1 (Table 2). *P. kernoviae* survived as peat-based inoculum exposed to compost for 1 day at
2 35°C but did not survive this treatment as cellophane or leaf inocula. There was no
3 survival of *P. kernoviae* after exposure to compost for 3 days at 35 °C or for 1 day at
4 40°C. No viability was detected in the leaf inoculum samples that were exposed to
5 compost to 35 or 45°C for 1 or 3 days, and then for a further 14 days at 20°C, i.e. all
6 values were zero (dormancy testing).

7 *P. ramorum*, as cellophane, leaf and peat-based inocula, survived in compost at
8 35°C after 5 days but was undetectable after 10 days (Table 2). There was no survival of
9 *P. ramorum* after exposure to compost at 40°C for 1 day. No viability was detected in the
10 leaf inoculum samples that were exposed to compost at 35 or 45°C for 3 days or to 45°C
11 for 1 day and then for a further 14 days at 20°C, i.e. all values were zero.

12 ITS1 RNA levels (Fig. 1) showed a clear decrease with increasing compost
13 temperature (to 35°C) and exposure time (to three days) for *P. ramorum* leaf and
14 cellophane inoculum, but not for peat inoculum. No ITS1 target was detected after one
15 day of incubation at 45°C in any of the three types of inoculum. At 35°C, results between
16 inoculum types were variable: cellophane and peat inoculum maintained a low level of
17 RNA after three days (Fig. 1a,c), whereas none was detected on leaves after three days
18 (Fig. 1b). Generally the level of ITS RNA showed the same trend as the culturing results
19 (Table 2), except cultures from leaves were positive after exposure to 35°C for three days
20 whereas no RNA was detected. Also, after a three-day exposure to compost at 18 °C, all
21 inoculum types produced positive cultures but no RNA was detected in leaf and peat
22 samples. In the flask experiments, RNA analysis was conducted on different samples to
23 those used for culturing, so some incongruity between the test results could be expected.

1 RNA results for inoculum samples exposed to compost for an additional 14 days at 20 °C
2 (dormancy testing) were all negative.

3 **Large-scale composting**

4 Mean and maximum temperatures of 58–64°C and 62–70°C respectively were achieved
5 in the centre of each of the four batches of shredded plant wastes at Trengwainton. The
6 minimum mean temperatures in the composting plant wastes were recorded in the lower
7 corners of the bay (15–44°C). Mean ambient air temperatures ranged from 6°C during the
8 fourth composting batch to 18°C during the second composting batch. Mean and
9 maximum temperatures in the centre of the three batches of composting bilberry waste at
10 Cannock Chase were lower than those recorded at Trengwainton (49–53°C and 58–60°C
11 respectively). The minimum mean temperatures in the composting waste were recorded
12 in the lower corners of the container (23–29°C). Mean ambient air temperatures were 11–
13 15°C. Standard deviations in temperature of the composting wastes were about 10% of
14 the mean values in each recording location at both sites.

15 The plant wastes used in the large-scale systems had higher moisture contents and
16 lower pH values and ash contents than the green waste composts used in the flasks.
17 During the 10-day composting periods, compost moisture declined and pH increased
18 slightly (Table 1). Mean O₂ and CO₂ concentrations in the composting plant wastes were
19 lower and higher respectively than those recorded in the flasks (Table 1). Ammonia
20 concentrations in woody plant waste composts at Trengwainton were only just detectable
21 with gas detector tubes ; no ammonia was detected in the composting bilberry wastes at
22 Cannock Chase (Table 1).

Post-composting, leaves that tested positive for *P. kernoviae* or *P. ramorum*, or bilberry stems that tested positive for *P. pseudosyringae* were only retrieved from the corners of the composting systems. All leaves retrieved from bags positioned in the surface, base, sides, back, front and centre of the composting bay, where mean compost temperatures were $\geq 41.9^{\circ}\text{C}$ over a 5-day period, tested negative for both *P. kernoviae* and *P. ramorum*. All bilberry stems retrieved from the same relative positions in the composting container, where mean compost temperatures were $\geq 37.3^{\circ}\text{C}$ over a 10-day period, tested negative for *P. pseudosyringae*. There were strong negative effects of mean compost temperature on the proportions of leaves that tested positive for *P. kernoviae* ($P = 0.001$, Fig. 2a) or *P. ramorum* ($P = 0.001$, Fig. 2b) or stems that tested positive for *P. pseudosyringae* ($P = 0.001$, Fig. 2c) post-composting. The fitted regressions in Fig. 2 have the following equation:

$$(1) \quad \ln(P/(1-P)) = bT + a + \ln(I_0)$$

where P is the proportion of positive *P. kernoviae* or *P. ramorum* leaves or *P. pseudosyringae* stems after composting (after exposure times of 5 or 10 days), T is the mean compost temperature, I_0 is the proportion of positive *P. kernoviae* or *P. ramorum* leaves or *P. pseudosyringae* stems in the day 0 samples, and a and b are the values of the model parameters (intercept and slope). The fitted values in Equation (1) for a , b and observed value for I_0 for exposures in composting plant wastes of 5 and 10 days of *P. kernoviae* and *P. ramorum* infected leaves and *P. pseudosyringae* infected stems are shown in Table 3.

The 10-day exposure in composting plant wastes resulted in a significantly ($P=0.003$) smaller proportion of leaves that tested positive for *P. kernoviae* than the 5-

1 day exposure. There was no evidence that the effect of mean compost temperature on the
2 proportion of leaves that tested positive for *P. kernoviae* was different between the
3 exposure times. However, only six observations (out of 46) showed some survival of *P.*
4 *kernoviae* after composting (four after 5 days and only two after 10 days). After 5-days
5 exposure to mean compost temperatures exceeding 32.8°C, there were no leaves that
6 tested positive for *P. kernoviae* (Fig. 2a).

7 Unlike for *P. kernoviae*, there was no overall difference between the two compost
8 exposure times, 5 and 10 days, in the proportion of leaves that tested positive for *P.*
9 *ramorum* post-composting in the large-scale system. However, there was a significant
10 interaction ($P = 0.002$) between the effects of compost temperature and exposure time,
11 with a more rapidly decreasing proportion of *P. ramorum* survival with increasing
12 temperature for leaves exposed for 10 days than for those exposed for 5 days. This is
13 shown by the greater negative slope for 10 days than for 5 days exposure in Table 3.
14 After exposure to compost for 5 days, there were no leaves that tested positive for *P.*
15 *ramorum* at compost temperatures of $\geq 41.9^{\circ}\text{C}$; the corresponding temperature for a 10-
16 day exposure was 31.8°C. The fitted survival probabilities of *P. kernoviae* and *P.*
17 *ramorum* on leaves, or *P. pseudosyringae* on stems, based on all the test data, indicate
18 that survival is still a small possibility, even at higher temperatures (Fig. 2). The
19 probabilities of obtaining a positive result for *P. kernoviae* or *P. ramorum* on leaves or *P.*
20 *pseudosyringae* on stems, exposed to compost for 5 days at 45°C or for 10 days at 35°C,
21 for the initial mean proportions of infected leaves or stems shown in Table 3, were less
22 than 3% (Table 4).

RT-Taqman results from the large scale composting experiment showed the same trend as culture results, with a logarithmic decline in ITS1 RNA in both species correlated with increasing temperature (Fig. 3). However, in both *P. kernoviae* and *P. ramorum* there were two important differences between ITS1 RNA measurements and viability as assessed by culturing. First, in the temperature range 41–45°C no viable cultures were found, but RNA was still detected at approximately 5% of the level observed at 16–20°C; second, at temperatures >53°C there was a higher level of ITS1 RNA than that recovered from leaf samples exposed to a temperature range of 41–45°C.

Discussion

The observation that *P. kernoviae* has a lower compost temperature tolerance than *P. ramorum* is in agreement with *in vitro* results of Turner et al. (2008) who showed that mycelium of *P. kernoviae* was killed after shorter exposure periods and at lower temperatures than mycelium of *P. ramorum*. Although *P. ramorum* was undetectable after a 1-day exposure to a constant flask compost temperature of 40°C, leaf inoculum survived a 5-day exposure to a mean temperature of 40.9°C in composting plant wastes in a large-scale system. These results are in broad agreement with those of Tooley et al. (2008) who reported that *P. ramorum*, as free chlamydospores and in rhododendron tissue, could survive a 4-day treatment at 35°C but not at 40°C. The pathogen survived a 2-day treatment at 40°C as chlamydospores but not in infected leaf tissue. Turner et al. (2008) found that wet heat treatment at 40°C was effective in killing *P. ramorum* mycelium after 15 min, at 42.5°C was effective in killing sporangia after 10 min, and at

1 45 °C was effective after 20 min in eradicating *P. kernoviae* and *P. ramorum* from leaves
2 inoculated with sporangia 24 h earlier. Swain et al. (2006) showed that *P. ramorum* could
3 not be isolated from infested leaves, wood chips and cankered stems after 24 h exposure
4 to dry heat at 40°C. Harnik et al. (2004) were able to re-isolate *P. ramorum* from
5 artificially inoculated California bay laurel (*Umbellularia californica*) leaves held at
6 55°C for up to 1 week, although such temperature tolerance has not been demonstrated
7 by other workers. Linderman & Davis (2006) found that *P. ramorum* survived as
8 chlamydospores for 12 months and as sporangia for 6 months in soil and various potting
9 mix components but could not be detected in infected rhododendron pieces after 1 month.

10 The majority of tests on other *Phytophthora* species, including *P. cactorum*, *P.*
11 *cinnamomi*, *P. infestans*, *P. megasperma*, and *P. nicotianae*, have shown that exposure to
12 a temperature of $\leq 50^{\circ}\text{C}$ for ≤ 7 days is sufficient to achieve eradication, as defined earlier
13 (Noble & Roberts, 2004; Noble et al. 2009). However, Noble et al. (2009) found survival
14 of oospores of *P. nicotianae* after exposure to compost at 54°C for 7 days, and Downer et
15 al. (2009) found survival of *P. cinnamomi* for up to 7 days in composting piles that
16 reached 60°C. Results here indicate that compost temperatures that are effective in
17 reducing inocula of *P. kernoviae* and *P. ramorum* to be below detectable limits, are also
18 effective for plant material infected with *P. pseudosyringae*.

19 Previous studies have shown RNA quantification as a reliable indicator of
20 viability in *Phytophthora* (Chimento et al., 2008). We have found that this is true only
21 within a certain temperature range, i.e. within which enzymatically driven RNA
22 degradation (both intra- and extracellular) can operate. However, at elevated
23 temperatures, RNA degradation no longer corresponded with *Phytophthora* viability and

1 ITS1 RNA levels increased above 53°C, compared to lower temperatures, even though no
2 viable *Phytophthora* were found in the cultures of the same samples. The most likely
3 explanation for this is that elevated temperatures prevented enzymatic degradation of
4 RNA. Juvonen et al. (2010) also observed this effect after heat treatment of
5 *Lactobacillus*, but it was not observed by Yaron et al. (2002) in *Eschericia coli* after heat
6 treatment. Many RNAses, e.g. RNaseA, are known to be thermally stable, but it could be
7 the case that the compost environment was not favourable for such thermostable enzymes
8 or that they were not present. It is also possible that heat shock temporarily raised
9 expression levels of ITS1 RNA prior to killing *Phytophthora* cells, therefore giving the
10 observed increased RNA level compared to lower incubation temperatures since heat-
11 shock has been shown to raise rRNA expression in other organisms (de Leon et al., 1997;
12 Carlson et al., 1999). A combination of chemical, physical and biological factors is likely
13 to influence the persistence of RNA in different environments. Therefore without prior
14 validation of non-cellular RNA in such conditions and their careful control during
15 experiments, RNA cannot reliably be used as an indicator of viability. It should be noted
16 that this may be true for any species, not just *Phytophthora*. This problem could be
17 overcome by measuring RNA only after a further brief incubation of a sample at
18 optimum temperature for the target organism, just long enough to allow degradation of
19 exogenous RNA from any dead organisms, however, this would negate the quantitative
20 and speed advantages of using RNA and RT-Taqman methods. Also, if the RNA
21 preservation was due to chemical conditions the extra incubation would still not degrade
22 RNA from dead cells. In complex environments such as the compost examined here, it
23 cannot be determined beforehand whether RNA preservation may occur.

1 The higher temperature tolerance of *P. ramorum* observed in the large-scale
2 composting system compared with that observed in the flasks may be due to differences
3 in the natural and artificially produced inoculum, and/or the substrates used, especially
4 peat, such as water potential. Turner et al. (2008) and Jennings (2008) found that
5 exposure to dry heat required longer and hotter treatments to achieve eradication of *P.*
6 *ramorum* than exposure to wet heat. This is consistent with the effect of compost
7 moisture on eradication temperatures of *Phytophthora nicotianae* and other pathogens
8 (Noble et al. 2009). Although the plant wastes used in the large-scale system had a higher
9 gravimetric moisture content than the flask composts, much of this was contained within
10 stems and leaves, rather than the freely available moisture applied to the mature compost
11 used in the flasks. The mature flask compost had a higher pH value than the plant wastes
12 used in the large-scale system (Table 1). Waste feedstocks have been found to exert an
13 effect on the temperatures needed to eradicate *P. nicotianae*, although the effect of
14 compost pH within the range 6–9.5 on pathogen eradication is usually minimal (Noble et
15 al. 2009). Even in the absence of elevated temperatures, composts, through biotic and
16 abiotic antagonism, are known to reduce the viability of the inocula of *Phytophthora*
17 species (Noble & Coventry, 2005) including *P. ramorum* (Fichtner et al. 2009).

18 Mean temperature values were used in the fitted regressions for compost
19 temperature and the proportion of *Phytophthora* positive leaves (Fig. 1). This was
20 justified in that the standard deviations in compost temperatures in sampling locations
21 were small due to the composting wastes first being allowed to rise in temperature before
22 leaves were inserted and the composting system being insulated against subsequent heat
23 loss. However, in situations where compost temperatures are more variable, other

parameters such as peak temperature or ‘day-degrees’ may be more appropriate for relating temperatures to pathogen survival characteristics (Noble & Roberts, 2004; Noble et al., 2009).

Temperatures above those required to reduce levels of *P. kernoviae*, *P. ramorum* and *P. pseudosyringae* to below detectable limits were achieved in the majority of the composting systems filled with shredded and wetted woody and leaf wastes from rhododendron and other shrubs and trees, or chopped bilberry. Survival of these pathogens was only detected in infected leaves positioned in bags in the cooler upper and lower corners of the bay. Survival risk was minimised by refilling the composting wastes into a second bay where the separated corner material was positioned into the centre and then achieved sanitizing temperatures on re-heating. By using the results of this work, the risk of survival of *P. kernoviae*, *P. ramorum* and *P. pseudosyringae* in a composting process can be estimated by monitoring time-temperature profiles in different locations, particularly in waste in the sides and corners of the system. This information can be used in decisions regarding the subsequent use, disposal or further treatment requirements of the composted waste. A well-managed and monitored composting system can therefore be considered to be an effective and phytosanitary treatment method for plant wastes infected with the *Phytophthora* species studied in this work.

Conclusions

1. A mean temperature of 35°C in flask or large-scale composts reduced *P. kernoviae* and *P. ramorum* to undetectable levels after exposures of 5 and 10 days respectively.

2. Although *P. ramorum* was undetectable after a 1-day exposure of different types of inoculum to green waste compost at 40°C in a flask, it survived in infected leaves exposed to a mean temperature of 40.9°C for 5 days in a large-scale composting system filled with shredded plant wastes.
3. No survival of *P. kernoviae* or *P. ramorum* was detected after exposure of infected leaves for 5 days to a mean temperature of ≥ 41.9 °C (32.8°C for *P. kernoviae*) or for 10 days at a mean temperature of 31.8°C (25.9°C for *P. pseudosyringae* on infected bilberry stems) in composting plant wastes a large-scale system.
4. The fitted probabilities of *P. ramorum* infected leaves being positive following exposure in a large-scale composting system for 5 days at 45°C or for 10 days at 35°C, with an initial 59.2% infected leaves, were less than 3%.
5. Compost temperatures achieved in insulated composting systems filled with shredded and wetted stem and leaf wastes of rhododendron, bilberry and other woody plants were above those needed to reduce levels of inoculum of *P. kernoviae* and *P. ramorum* in infected leaves and *P. pseudosyringae* in infected stems to below detectable limits, except in the upper and lower corners of the systems. Pathogen survival risk can be minimised by ensuring the corner material is filled into the centre of the composting system after re-filling, and by monitoring compost temperatures, particularly in the sides and corners of the system.
6. RNA methods of quantifying viability are not reliable in environments that favour RNA preservation.

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Table 1. Properties of green waste composts used in flasks and of wastes used in large scale composting systems, before wetting, and at the start and at the end of the temperature incubation experiments, and gaseous conditions in the materials during incubation. Each value is the mean of at least three replicate samples.

Compost or waste	Moisture (% w/w)			pH		Ash (% DM)		O ₂	CO ₂	NH ₃
	before	start	end	start	end	start	end	(% v/v)	(% v/v)	(mg m ⁻³)
Green waste compost	32	41	38	7.6	7.4	76.5	78.0	17.4	2.8	0.4
Woody plant wastes	48	63	56	6.1	6.5	2.6	3.7	13.1	4.2	0.2
Bilberry waste	45	64	61	5.1	5.9	3.8	5.2	12.7	4.6	0

Table 2. Viability of *Phytophthora kernoviae* and *P. ramorum* in different inoculum media following different temperatures and exposure times in flask composts, determined by semi-selective culturing. Each flask contained three samples of cellophane and leaf inoculum and a sample of peat-based inoculum. Each value is the mean of six flasks, except for 5 and 10 day, and 40°C treatments ($n = 2$) and 3 days at 45 °C treatments ($n = 4$).

Time (days)	Medium ^a	Temperature (°C)							
		18		35		40		45	
		<i>kernoviae</i>	<i>ramorum</i>	<i>kernoviae</i>	<i>ramorum</i>	<i>kernoviae</i>	<i>ramorum</i>	<i>kernoviae</i>	<i>ramorum</i>
1	cello. (%)	67	100	0	72	0	0	0	0
	leaf (%)	100	100	0	78	0	0	0	0
	peat (cfu)	1.3×10^4	1.8×10^4	0	1.4×10^4	0	0	0	0
3	cello. (%)	72	94	0	6	0	0	0	0
	leaf (%)	100	83	0	39	0	0	0	0
	peat (cfu)	3.1×10^4	9.3×10^3	0	2.4×10^3	0	0	0	0
5	cello. (%)	0	100	0	17	- ^b	-	-	-
	leaf (%)	67	100	0	17	-	-	-	-
	peat (cfu)	1.0×10^5	3.7×10^4	0	4.2×10^3	-	-	-	-
10	cello. (%)	0	100	0	0	-	-	-	-
	leaf (%)	83	100	0	0	-	-	-	-
	peat (cfu)	7.1×10^3	1.0×10^5	0	0	-	-	-	-

^a mean values for cellophane and leaf inoculum media are percentages of positive samples and mean values for peat-based inoculum are cfu per g.

^b dash indicates treatment not tested

Table 3. Fitted slope and intercept, on the logit scale, for the regressions between mean compost temperature and the proportion of positive leaves with *Phytophthora kernoviae* or *P. ramorum* or positive stems with *P. pseudosyringae* after exposure to composting plant wastes for 5 or 10 days in large-scale systems. I_0 is the mean proportion of positive leaves or stems at day 0.

<i>Phytophthora</i> species	Exposure (days)	Slope (s.e.)	Intercept (s.e.)	I_0
<i>P. kernoviae</i>	5	-0.16 (0.04)	3.83 (0.84)	0.084
<i>P. kernoviae</i>	10	-0.16 (0.05)	2.46 (1.11)	0.084
<i>P. ramorum</i>	5	-0.10 (0.02)	1.67 (0.63)	0.592
<i>P. ramorum</i>	10	-0.27 (0.06)	5.10 (1.26)	0.592
<i>P. pseudosyringae</i>	10	-0.27 (0.08)	4.24 (1.27)	0.404

Table 4. Fitted probability and 95% confidence intervals of obtaining *Phytophthora kernoviae* and *P. ramorum* positive leaves or *P. pseudosyringae* after exposure to composting plant wastes for five days at 45°C or ten days at 35°C in large-scale systems, for the mean proportions of positive leaves at the start (I_0) in Table 3.

<i>Phytophthora</i> species	Fitted probability (and 95% confidence intervals) (%)	
	5 days at 45°C	10 days at 35°C
<i>P. kernoviae</i>	0.29 (0.05, 1.59)	0.40 (0.09, 1.79)
<i>P. ramorum</i>	2.73 (1.14, 6.39)	0.76 (0.13, 4.27)
<i>P. pseudosyringae</i>	-	0.22 (0.02, 1.90)

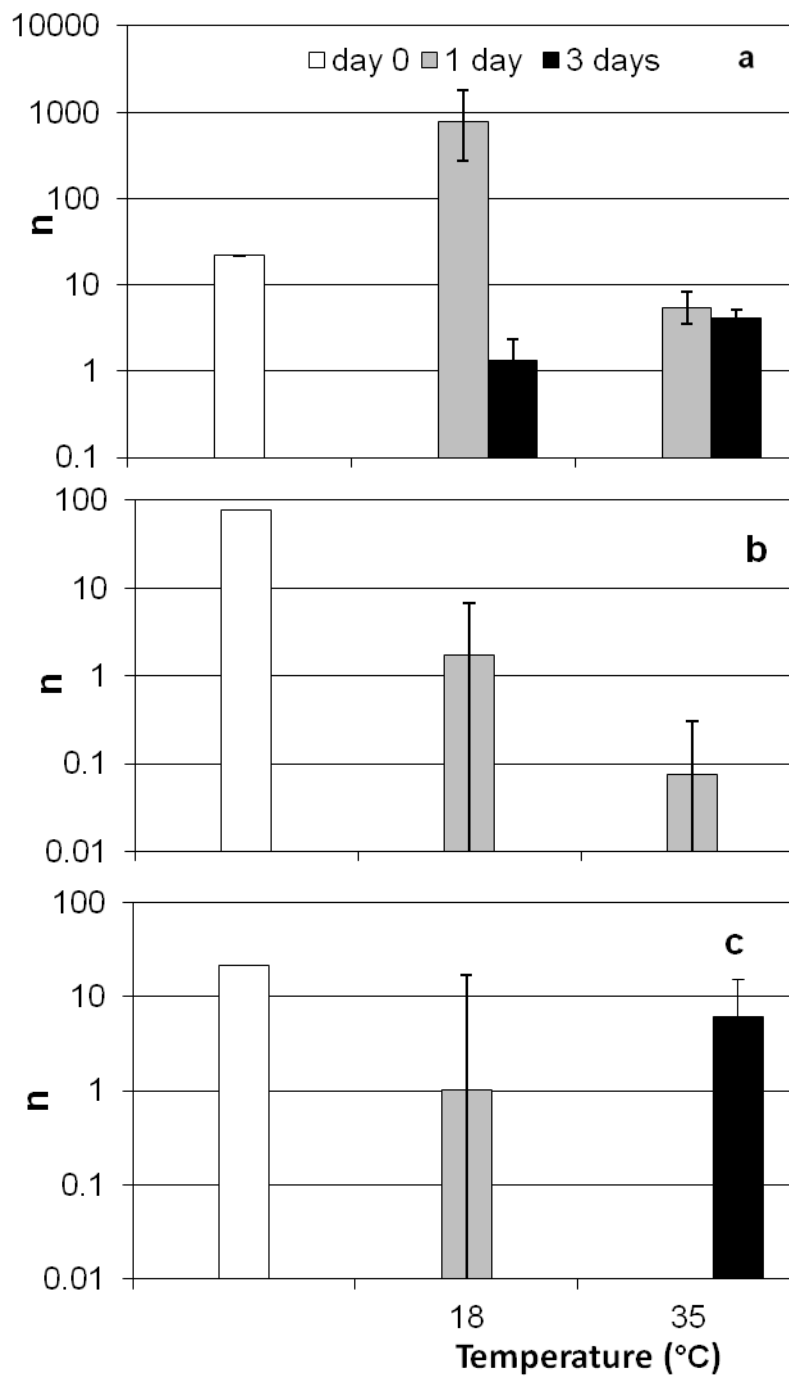


Figure 1. Mean relative amount of ITS1 RNA, as measured by RT Taqman PCR, in (a) cellophane (b) leaf and (c) peat samples of *P. ramorum* inoculum following different temperatures and exposure times in flask composts. No RNA was detected in samples exposed to compost at 45°C. Each value is the mean of two replicate flasks.

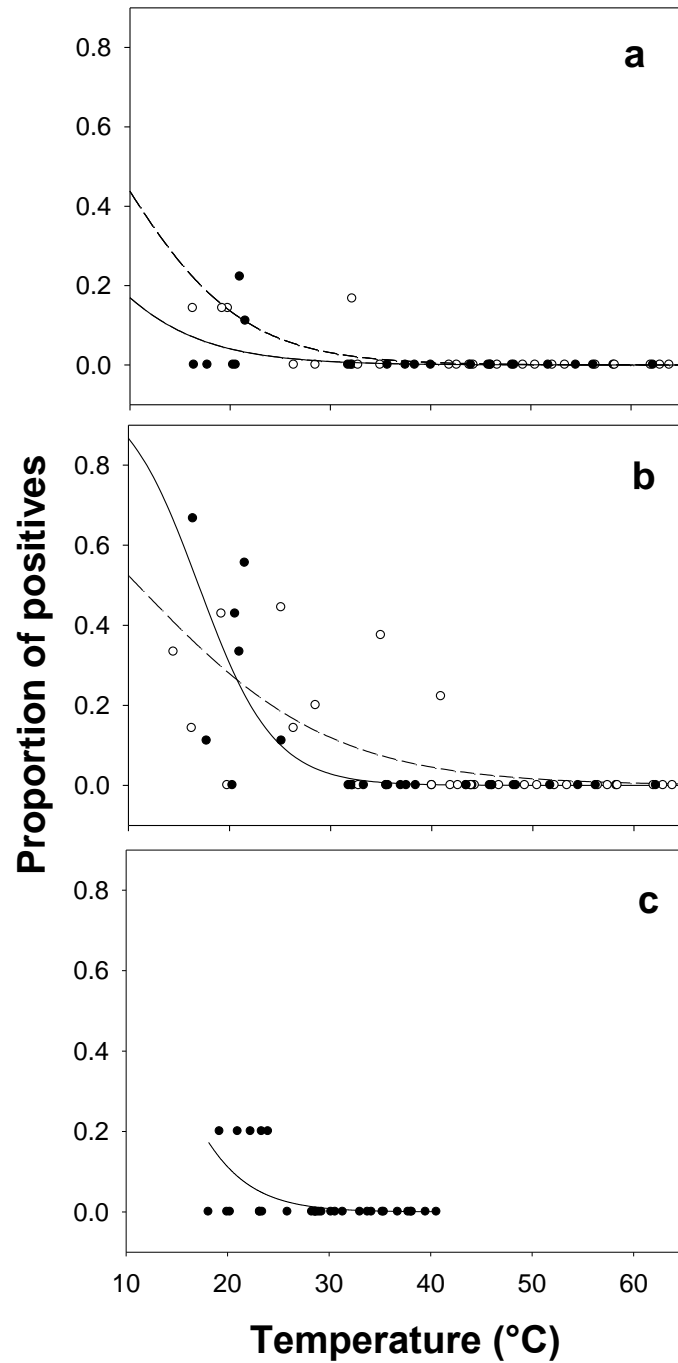


Figure 2. Fitted regressions for mean compost temperature and proportion of positive leaves of (a) *Phytophthora kernoviae* (b) *P. ramorum* and (c) *P. pseudosyringae* after 5 days (\circ and dashed line) and 10 days (\bullet and solid line) in composting plant wastes in a large-scale system. The values in the regression equation (1) are shown in Table 3.

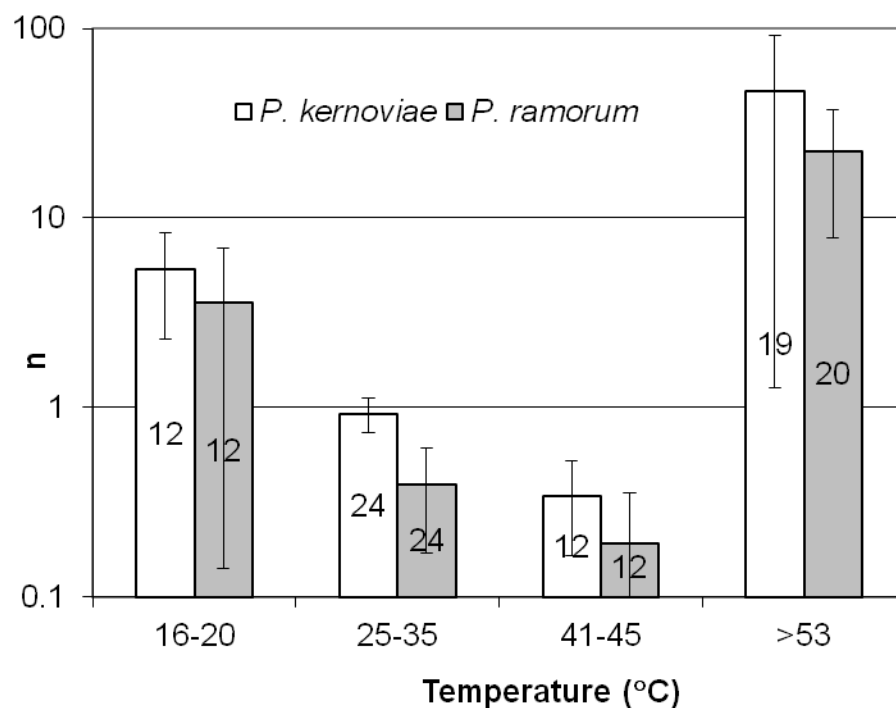


Figure 3. Mean relative amount of ITS1 RNA, as measured by RT Taqman PCR, in leaves retrieved from four temperature ranges in a large-scale composting system. Error bars are the 95% confidence limits of the measurements at each temperature. The value in each bar is the number of samples in each temperature range.